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Exposure to indoor air contaminants in school buildings with and without reported indoor air quality problems



Camilla Vornanen-Winqvist^{a,*}, Kati Järvi^a, Maria A. Andersson^a, Caroline Duchaine^{b,c}, Valérie Létourneau^c, Orsolya Kedves^d, László Kredics^d, Raimo Mikkola^a, Jarek Kurnitski^{a,e}, Heidi Salonen^a

^a Aalto University, Department of Civil Engineering, PO Box 12100, FI-00076 Aalto, Finland

^b Université Laval, Département de Biochimie, de Microbiologie et de Bio-informatique, 2325, rue de l'Université, Québec G1V 0A6, Canada

^c Centre de Recherche de l'Institut Universitaire de Cardiologie et de Pneumologie de Québec, 2725, chemin Sainte-Foy, Québec G1V 4G5, Canada

^d University of Szeged, Faculty of Science and Informatics, Department of Microbiology, Közép fasor 52, H-6726 Szeged, Hungary

^e Tallinn University of Technology, Department of Civil Engineering and Architecture, Ehitajate tee 5, 19086 Tallinn, Estonia

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ABSTRACT

Reported indoor air quality (IAQ) complaints are common even in relatively new or renovated school buildings in Finland. However, detecting the causes for complaints with commonly used indoor air measurements is difficult. This study presents data on perceived and measured IAQ in six comprehensive school buildings in Finland. The aim of this study was to discover the possible differences of perceived and measured IAQ between schools with reported IAQ complaints and schools without reported IAQ complaints. The initial categorisation of schools with ('problematic schools') and without ('comparison schools') complaints was ensured via a validated indoor climate survey and a recently developed online questionnaire, which were completed by 186 teachers and 1268 students from the six schools. IAQ measurements of physical parameters, gaseous pollutants, particulate matter and bioaerosols were conducted in four problematic school buildings (26 classrooms) and two comparison school buildings (12 classrooms). Using air sampling as well as exhaust air filters and classroom settled dust to detect the presence of elevated concentrations of airborne cultivable microbes and pathogenic, toxigenic and mycoparasitic *Trichoderma* strains were the most indicative methods in distinguishing problematic schools from comparison schools. Other IAQ-related measurements did not detect clear differences between problematic and comparison schools, as the concentration levels were very low. The results indicate that the complaints reported by occupants could have been related to excess moisture or mould problems that had not been found or repaired. Ventilation pressure condition investigations and simultaneous exhaust and supply air filter dust culture should be addressed precisely in future studies.

1. Introduction

Schools are places with a high level of activity and population density of children, where different pollutants from both indoor and outdoor sources may be introduced and persist for a long time (Chithra and Nagendra 2018). Additionally, school buildings are often characterised by infrequent interventions for building maintenance and environmental remediation (Theodosiou and Ordoumpozanis 2008, Ruggieri et al. 2019).

Indoor air quality (IAQ) in school buildings is characterised by various pollutants, such as volatile organic compounds (VOCs), aldehydes, particulate matter (PM_{2.5} and PM₁₀), fungi and bacteria

(Madureira et al. 2009, Salonen et al. 2015, Madureira et al. 2012). Relationships between specific indoor exposures—even at low levels—and the development or exacerbation of respiratory symptoms and asthma have been shown among allergic persons, in particular, and individuals who have not been sensitised before (Madureira et al. 2009, Madureira et al. 2015, Kanchongkittiphon et al. 2015). In addition to occurring indoor air pollutants, indoor environmental conditions and contributors such as high temperature, low relative humidity and the ventilation quality strongly affect IAQ and may cause or contribute to health effects and affect the comfort in indoor environment (Turunen et al. 2014, Mendell and Mirer 2009, Lan et al. 2011, Torresin et al. 2018).

* Corresponding author.

E-mail address: camilla.vornanen@aalto.fi (C. Vornanen-Winqvist).

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Children are more susceptible to the effects of air contaminants than adults, due to their immature immune and respiratory system, inferior body mass index and breathing pattern (Yassin and Pillai 2018, Chithra and Nagendra 2018, WHO 2006, Mendell and Heath 2005, Faustman et al. 2000). As a result of the time spent in schools, indoor environmental conditions are among the major contributors of total exposure for children to various air pollutants (Lai et al. 2015, Sheehan et al. 2012, Jacobs et al. 2013, Yassin and Pillai 2018, Bennett et al. 2019, Salthammer et al. 2016).

In addition to the increased chance of long- and short-term health problems for students and staff, air pollutants and poor IAQ reduce children's school attendance, academic performance (Mendell and Heath 2005, Annesi-Maesano et al. 2013) and comfort level (Chithra and Nagendra 2018). They also reduce teachers' productivity and increase sick leaves (Ervasti et al. 2012, Chithra and Nagendra 2018). Thus, the characterisation of different indoor air pollutants and their effects on measured and perceived IAQ in school environments is of critical public health concern, given the potential long-term adverse consequences of the exposures. However, the available IAQ data from school buildings regarding the characterisation of simultaneously occurred pollutants and situation-related perceived disadvantage and symptom mapping are limited. For example, studies regarding IAQ in Finnish comprehensive schools typically focused on microbial investigations (Peitzsch et al. 2012, Haverinen-Shaughnessy et al. 2012, Meklin et al. 2005, Husman et al. 2002), and none have collected wide, simultaneous data of perceived and measured IAQ in the Helsinki metropolitan area.

The aim of the present study was to conduct a comprehensive characterisation of the IAQ in school buildings with and without reported IAQ complaints and detect measurable differences between the buildings. Occupants' perceptions were recorded to ensure the initial categorisation of buildings with and without perceived IAQ problems. Characterisation was done by measuring ventilation, physical parameters and the above-mentioned main air pollutants: gaseous pollutants, particulate matter, and indoor microbiota, focusing on toxigenic, mycoparasitic and potentially pathogenic moulds.

2. Material and methods

2.1. Study design and school characteristics

The present study was conducted as part of Finnish research projects (EURA, TOXICPM and Indoor Air Police) investigating the chemical, physical and microbiological IAQ in comprehensive school buildings. The study was carried out at six public schools in the Helsinki metropolitan area—two schools in Helsinki (P4 and C6) and four schools in Vantaa (P1, P2, P3 and C5). The suitable buildings were selected with the help of indoor air specialists from Real Estate Centre of Vantaa and Built Assets Management of City of Helsinki, as officials responsible for building and IAQ problem management and for ensuring proper

working environment for the employees and students. Four of the selected buildings were designated as 'problematic schools' (P1–P4) and two were comparison schools (C5 and C6) based on the occupants' IAQ complaints or lack of complaints, respectively, received by the indoor air specialists and work health personnel.

The indoor air specialists and the experts they consult from several fields had the best knowledge of the buildings under study since they had controlled all the information related to, for example, maintenance and repair actions, IAQ investigations, building technology data and problem reports, as well as IAQ questionnaire and work place investigation reports implemented by occupational health. The complaints from the schools were reported directly or via occupational health care personnel to them. Therefore, the information about the buildings' perceived and measured problems received from the indoor air specialists of the two cities was accurate and reliable for this study and was also confirmed by each school's principal.

Complaints had been received by the indoor air specialists and occupational health personnel from the problematic schools for at least one to two years. They included, for example, perceptions of poor, stuffy and warm air, dust, bad odours, headaches, nausea, skin symptoms, and respiratory symptoms, such as dry cough, hoarseness, stuffy nose or dry throat. Seasonal and psychosocial factors affect the perceptions of symptoms and discomfort, but in these buildings the variety of symptoms, inconvenience, and their continuation over a long period of time suggested other causes for the symptoms. According to the indoor air specialists, several investigations and ventilation, sealing and structural repairs had been conducted in these buildings over the past few years; however, complaints continued and a clear reason for occupants' perceptions had not been found. Thus, the selected problematic schools were considered buildings with prolonged, unsolved IAQ problems. Comparison schools were selected because there were no IAQ-related complaints and no existing repair needs, or work health issues were known by the school principals or indoor air specialists. Before the study began, school personnel and parents were informed about the aims of this study and whether their school was categorized as problematic or was a comparison school.

The number of studied buildings was determined by the number of suitable schools that were available for the study and by project resources. Since the circumstances in the problematic schools were the focus of this study, the lower number of comparison schools was accepted. In addition to comparison schools, the available decrees on indoor conditions and recommended limit values provided reference information of good or acceptable indoor conditions, however, not replacing the valuable information from real school environments.

All schools were in urban population centres approximately 20–50 m from main traffic routes, but not in direct proximity of heavy traffic. School C6 was the nearest building to busy city routes. Sparse forest cover or plants existed in all the school areas. The study was conducted between October 2016 and March 2017, that is, during the winter season. Meteorological conditions of the various schools during

Table 1
Information on the studied schools.

School	Approx. no. of students	No. of employees	Year of construction	Year of renovation	Type of ventilation	Measurement period	Median outdoor T (°C) *	Median outdoor RH(%)*
P1	840	79	2006	–	mv	7. – 21.11.2016	–1.4	93
P2	840	75	2006	–	mv	21.11. – 7.12.2016	–0.2	91
P3	680	56	1968, extension 2002	2003–2005	mv	30.1. – 13.2.2017	–2.4	89
P4	700	74	2009	–	mv (classroom 6f), nv (6a–6e)	6. – 20.3.2017	0.8	88
C5	570	68	1980s, extension 2013–2014	–	mv	24.10. – 7.11.2016	1.2	85
C6	800	40	1955	2010–2012	mv	16. – 30.1.2017	0.2	94

P: problematic building (reported indoor air complaints); C: comparison building (no reported indoor air complaints); mv: mechanical supply and extract ventilation; nv: fan-assisted natural ventilation; *outdoor temperature (T) and relative humidity (RH) during the measurement period.

the measurement periods were comparable (median temperature from -1.4°C to 1.2°C and median relative humidity from 85% to 94%), as shown in Table 1, therefore not causing bias in the result analysis. Primary and upper comprehensive school education was given in the buildings, mainly from grades one to nine. Information on the studied schools is presented in Table 1.

From each school, six classrooms (a–f, later combined with the number of the school) were selected for the study by the school principals and personnel. In school P2, however, two additional classrooms (2 g and 2 h) were studied. The studied classrooms presented a comprehensive sample of indoor spaces from each building and could be studied simultaneously within the project's device and time resources. Classrooms were selected from different floors and sections of the schools. Locations of the studied classrooms in the layouts of the school buildings are presented in Fig. S1 in the supplementary material.

Principals were asked to select classrooms that were the most occupied during the school week and that the students taught in the classrooms would be from grades seven to nine (upper comprehensive school) to be mature enough to answer the online questionnaire. In addition, most of the selected classrooms complied with the request that any special emitting materials (arts, physics and chemistry) would not be used in the classrooms, with few exceptions, which have been noted in result analyses.

In problematic schools, the classrooms associated with the most severe reported indoor air problems were selected. The principals made these selections because they had the best viewpoint of the school's overall circumstances, perceived problems, previously conducted investigations and user questionnaires, and reports from the occupational and school healthcare in use. In all problematic schools, the escalated situation of prolonged and numerous complaints of perceived and unresolved IAQ problems negatively affected the working environment. Therefore, the personnel and principals had a strong will to contribute in studies providing new information about such difficult cases and tended to select the most problematic classrooms for this study. Association between school's IAQ problems and student-perceived social climate and the psychosocial atmosphere of the work community have been shown in previous studies (Finell et al. 2018, Lahtinen et al. 2002).

In the comparison schools, principals selected some classrooms in which some IAQ-related discomfort, such as too warm or stuffy air, had occurred. This classroom selection might have weakened the distinction between problematic and comparison schools. However, problems in comparison schools were not considered as frequent or as severe as in the problematic schools, thus the classrooms presented a 'normal' school environment at an acceptable level.

Although surface cleaning was done daily in those classrooms involved in the study, the daily classroom cleaning was not performed for four days before the measurement period. Since measurements were made on Mondays, classrooms were not to be cleaned after Thursday. In the 'walk-through' inspection prior to the sampling, the classrooms appeared clean and no visible sand, dust or other dirt could be seen on the floors, although mice excrement was observed in classrooms 3a and 3b. Dust on upper surfaces was common in all studied schools.

Surface materials in all the classrooms included plastic flooring as well as painted walls and ceilings. There were no animals or visible moisture or mould inside the classrooms. Visible sources for man-made vitreous fibres could be seen in the ceiling structures of classrooms 6a and 3c. Air cleaners were in use in classrooms 1e, 2b, 3c and 4a, but, according to teachers, were kept off during the measurement period.

2.2. Indoor climate questionnaires

2.2.1. Questionnaire for the employees

Employees' indoor air-related discomfort and symptoms, which had occurred during the past three months in the work environment, were recorded with the validated and frequently used indoor climate survey

of the Finnish Institute of Occupational Health (FIOH) in January 2017. The survey is based on the Örebro (MM40) Indoor Climate Questionnaire (Andersson 1998). Occupants working throughout the schools were asked to participate during the two weeks allotted for responses. The schools' principals were responsible for delivering the questionnaires, while the FIOH collected and reported the answers.

Results presented in this study were based on the school-specific indoor climate survey reports received from the FIOH. In the reports, results (percentage of respondents reporting an occurrence of a certain symptom or work environment factor) were compared to reference data collected from 122 office and healthcare workplaces in Finland (Reijula and Sundman-Digert 2004, Helgren 2008), presenting IAQ in a 'normal' working environment. However, it must be noted that schools have some differences from those environments in, for example, noisiness and ventilation sufficiency. Additionally, the psychosocial load or other factors related to work environment or personal conditions can worsen the perceptions of IAQ (Runeson-Broberg and Norbäck 2013, Lahtinen et al. 2004); therefore, the questionnaire surveys work-related stress (i.e., tasks, responsibilities and office hierarchy) and allergic illnesses (Andersson 1998, Reijula and Sundman-Digert 2004).

2.2.2. Online questionnaire for the students and teachers

Symptoms and perceived IAQ were recorded during the two-week measurement period in the studied classrooms using a simple and short online questionnaire developed in the related research project, 'Indoor Air Police' (Järvi et al. 2018). The aim of the questionnaire development was to create an easy tool for time- and space-specific reporting of the occurring problems and to involve students as an important group of school occupants. Questions were selected based on previous questionnaires on IAQ, for example, the Örebro MM-60 and MM-80 (Andersson 1998). Both teachers and students from grades seven to nine responded the questionnaire following each class via a personal link by using a cell phone or a tablet computer. Paper questionnaires were available in case of technical problems. Questions addressed perceived symptoms and comfort during the previous class with simple answer options, such as 'no', 'little', and 'a lot'. When responding for the first time, background information including personal health (e.g., perceived health, allergic diseases and smoking) and home environment factors (e.g., pets, type of accommodation and moisture damage at home) were covered as well. Statistical differences between problematic and comparison schools and between classrooms were studied from the collected responses. This study was based solely on self-reports, and no objective measurements of occupants' health were conducted.

2.3. Sampling and instrumentations

Data acquisition consisted of one-minute samplings and long-term measurements for up to two weeks, as well as recordings of perceptions both on a real-time basis and retrospectively. The filter and settled dust and endotoxin samples and the two-week monitoring represented long-term accumulation and exposure. The aim was to form an encompassing outlook of multiple parameters in the studied schools with the available resources. Altogether, 14 different parameters were measured. Measured factors, measurement methods, devices and their accuracy, measurement locations and times are presented in Table S1 in the supplementary material.

A schematic diagram of one classroom and its measurement locations are presented in Fig. 1. All classrooms were shaped and furnished differently, but the measurement set-up was analogous in each.

2.3.1. Ventilation and physical parameters

Differential pressures (dPa) across the building envelope were measured in each classroom with SWEMA 3000md device (Swema, Stockholm, Sweden) as 60-second average measurements to observe the current pressure conditions in the rooms. Measurements were

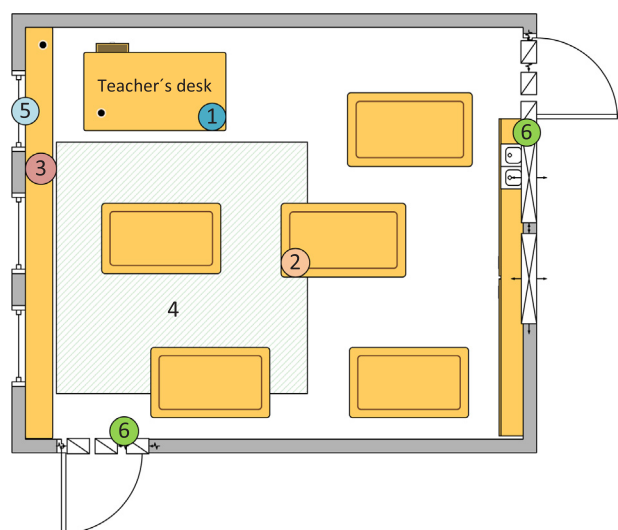


Fig. 1. Measurements in a classroom. (1) temperature, relative humidity and carbon dioxide measurement, (2) volatile organic compound sampling, (3) sampling of other gaseous pollutants and particulate matter, (4) endotoxin sampling area, (5) differential pressure measurement through window and (6) air flow rate measurement from the supply and extract air diffuser. In addition, airborne cultivable microorganisms were collected from air near the classroom walls, and in some classrooms, settled dust was wiped from the available surfaces near walls or above door height.

conducted with a plastic tube and a copper core placed outside of a closed window. Supply and extract air flow rates were evaluated as real-time readings from the ventilation terminal units in the classrooms where reliable measurements were possible during the visit.

Temperature (T), relative humidity (RH) and carbon dioxide (CO_2) concentrations were recorded as continuous measurements for a two-week period in each classroom with Rotronic CL11 meter (Rotronic AG, Bassersdorf, Zürich, Switzerland), see Fig. S2 in the [supplementary material](#). Measurement devices were placed on the teacher's desk at the front of the room, away from the teacher's breathing zone when seated and as close to the horizontal central area of the room as possible.

T and RH were compared to simultaneous regional outdoor conditions. The outdoor data were taken from the open access weather statistics of the [Finnish Meteorological Institute \(2019\)](#), measured at weather stations located approximately 1.5–10 km from the schools (10.0 km from P1, 8.2 km from P2, 4.5 km from P3, 1.5 km from P4, 4.5 km from C5 and 6.0 km from C6). In the database, data are available as 10-minutely values. Outdoor medians were calculated from the data collected during the corresponding school measurement period.

2.3.2. Gaseous pollutants and particulate matter

VOC sampling and analysis were carried out in each classroom according to the ISO 16000-6 standard (ISO 16000-6:2011). Duplicate air samples were taken from the central area of an empty and closed room in the main working zone at a height of 1.5 m. Samples of 8 dm³ were collected in stainless steel tubes (Markes International Ltd., Llantrisant, UK) packed with Tenax TA (60/80 mesh) and Tenax TA-Carbograph 5TD at a flow rate of 0.20 dm³/min using GilAir Plus air sampling pumps (Sensidyne, St. Petersburg, FL, USA). Analyses were conducted at Aalto University or the FIOH. Sampling of VOCs is shown in Fig. S3 in the [supplementary material](#).

Formaldehyde concentrations of indoor air were evaluated in classrooms 1e, 2g, 3a, 4d, 5b and 6d using a formaldehyde meter FM-801 (Graywolf Sensing Solution, Sheldon, IA, USA) for a two-week period. Carbon monoxide (CO), hydrogen sulphide, arsine, ammonia and ozone were measured in classrooms 1e, 2g, 3a, 4d and 5b as continuous two- to four-day measurements with TG-501 and IQ-610 probes and stored in a palm computer and online service (Graywolf Sensing

Solution, Sheldon, IA, USA).

Fine particulate matter ($\text{PM}_{2.5}$) was measured in classrooms 1e, 2g, 3a and 5b, and coarse particulate matter (PM_{10}) was measured in classrooms 1e, 2g, 5b and 6d for a two-week period using an MIE pDR-1500 nephelometer equipped with a $\text{PM}_{2.5}$ or PM_{10} size-selective inlet cyclone (Thermo Fisher Scientific, Franklin, MA, USA).

The measurement devices were placed in the back or front of the room at a height of 1.5 m and as close to the central area as possible. Measurement units are shown in Fig. S4 in the [supplementary material](#).

2.3.3. Endotoxin loads in settled surface floor dust

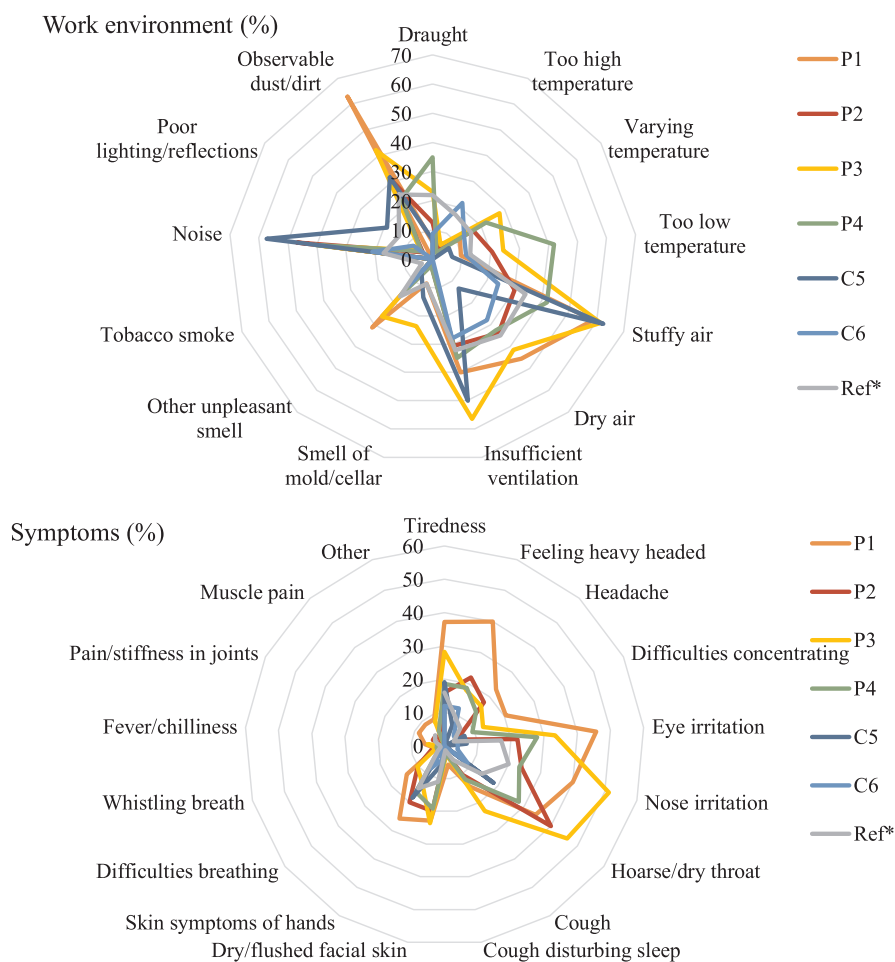
The endotoxin samples were collected from the classroom floors using a vacuum cleaner (Volta Equipt or Electrolux ZB5003, AB Electrolux, Stockholm, Sweden) with disposable wipers as filters (Kimwipes, Kimtech Science Wipers, Merck KGaA, Darmstadt, Germany). A total of 36 samples were collected from six schools (six classrooms in each school) representing long-term accumulation and exposure. Before sampling, the inside and outside of the vacuum cleaner nozzle were disinfected with 70% ethanol. Dust was vacuumed in an empty classroom during a regular school day for five minutes and from an approximately 10 m² area near children's desks or working areas (see Fig. 1). After sampling, filters were put into clean plastic bags (Minigrip, Amer Plast, Tampere, Finland) and frozen (−20 °C) (Duquenne et al. 2013, Letourneau et al. 2009, Spaan et al. 2008, 2007) until analysis.

The samples were analysed at the Centre de Recherche de l'Institut Universitaire de Cardiologie et de Pneumologie de Québec, Canada, using a Kinetic-QCL™ Kinetic Chromogenic LAL Assay (Lonza, Walkersville, MD, USA). The loads of endotoxins in floor dust were then expressed as EU/m². Briefly, endotoxins were extracted by vortexing (maximum speed, Multi-Pulse Vortexer, Glas-Col, Terre Haute, IN, USA) filters for 1 h at 4 °C in 30 mL of sterile pyrogen-free salt solution (0.9% NaCl) supplemented with 0.025% Tween 20. Solutions were then centrifuged at 500g for 10 min at 4 °C, and supernatants were used for endotoxin quantification. Inhibition/enhancement tests were performed, and the detection limit of the method was 0.47 EU/m². The method's efficiency, reproducibility and sensitivity for endotoxin quantification have been evaluated in previous studies (Duquenne et al. 2013, Letourneau et al. 2009, Spaan et al. 2007, Spaan et al. 2008).

2.3.4. Airborne cultivable microorganisms

Airborne cultivable microorganisms were monitored at indoor (two samples per each classroom, 12 samples per each school) and outdoor locations (two samples per each school) with a RCS® High Flow Touch Microbial Air Sampler (Merck KGaA, Darmstadt, Germany, Ser. No. 41877) and agar media contained in a flexible foil (HYCON®, Merck KGaA, Darmstadt, Germany), see Fig. S5 in the [supplementary material](#). The sampling volume used was 0.10 m³.

Tryptic Soy Agar (TSA, with no added fungicide) strips were used to identify the total cultivable microorganisms (TC, Total count (941105), HYCON®), and Rose Bengal Agar strips with streptomycin and chloramphenicol were used to selectively isolate and enumerate cultivable yeasts and moulds (YM, Yeasts and moulds (941196), HYCON®) according to the manufacturer's guidelines. Agar media blanks were taken into the field but not opened (Bartlett et al. 2004). All samples were collected within a 30 min time period in empty classrooms during regular school hours and in between normal room activities to reflect the building-related conditions that occupants are exposed to during the school day. Measurements were taken at a height of 1.0 m above floor or ground level, and by walking around the classroom or outdoors at the front area of the school facade. Strips were then incubated at two different temperatures (30 °C for total cultivable microorganisms and 21 °C for cultivable yeasts and moulds) prior to counting the visible colonies (after four days for TC, see Fig. S6 in the [supplementary material](#), and after seven days for YM).



*Reference values (Reijula and Sundman-Digert 2004)

Fig. 2. Perceived work environment and symptoms according to the FIOH's indoor climate survey.

2.3.5. Characterisation of the cultivable moulds in indoor airborne and settled indoor dust with special focus on toxigenic *Trichoderma* strains

Moulds in indoor airborne dust were extracted from the exhaust air filters of air handling units serving the building sections where the studied classrooms were located. Airborne moulds captured by the air filters (collecting a sample of filter dust to be cultivated is shown in Fig. S7 in the supplementary material) were cultivated directly on malt extract agar (MEA) inoculated by shaking the filters above the agar plates (malt extract 15 g; Sharlab, Barcelona, Spain; agar 12 g; Amresco, Solon, Ohio, USA, in 500 mL of H₂O). Settled dust was also collected by clean plastic bags (Minigrip, Amerplast, Tampere, Finland) from surfaces 1.5–2.0 m above the floor level of the classrooms under study in schools P3, P4 and C6, and cultivated on MEA.

Fungal colonies representing dominant morphotypes after four weeks of incubation were characterised and identified to the genus or species level by macroscopic and microscopic observations. Colonies of cultivable moulds were screened for toxicity. Sampling and characterisation methods, as well as toxicity screening, were described by Vornanen-Winqvist et al. (2018a). Attention was focused especially on the diversity of toxigenic *Trichoderma* strains, potentially pathogenic versus non-pathogenic *Rhizopus* strains, toxigenic *Fusarium* and toxigenic versus non-toxicogenic *Chaetomium*-like strains.

Fungal colonies were grouped into nine morphotypes, based on their morphology on MEA, ability to grow at 37 °C, light microscopy results for conidia and conidiophores, mycoparasitic growth and responses to the two toxicity assays: the boar sperm motility inhibition

assay and the inhibition of cell proliferation assay. The isolates were compared to the reference strains from the HAMBI culture collection (Microbial Domain Biological Resource Centre HAMBI, University of Helsinki, Finland) or identified according to the process described by Samson et al. (2004). Representatives of the morphotypes of toxigenic *Trichoderma* and *Chaetomium* strains were further identified by sequence analysis of the ribosomal RNA gene cluster's internal transcribed spacer region and/or a fragment of the gene encoding for translation elongation factor 1 alpha (*tef1*) (Castagnoli et al. 2018).

The biocide resistance of the isolated *Trichoderma* strains was tested with polyhexamethylene biguanide (PHMB, 30 µg mL⁻¹) and borax (600 µg mL⁻¹) MEA plates (15 g malt extract and 12 g agar per 500 mL of H₂O). PHMB-containing MEA plates were prepared by adding 15 mg of PHMB to 500 mL of the water used to prepare the 500 mL of the MEA medium. Borax-containing plates were prepared via the process described by Mikkola et al. (2015). The plates were inoculated with about 5 mg of spores, incubated for two weeks at 22 °C and kept in daylight to optimise sporulation. Growth and sporulation were inspected after one and three weeks. Three tests were repeated per strain, and MEA control plates were inoculated in parallel.

2.4. Statistical analysis

Statistical parameters were calculated with Microsoft Office Excel. The statistical analyses of the differences between the online questionnaire results were performed by cross tabulation, Pearson χ^2 -test

and Fisher's exact test (SPSS-25.0 for Windows, SPSS, Inc., Chicago, IL, USA).

3. Results and discussion

3.1. Indoor climate questionnaires

3.1.1. Questionnaire for the employees

The number of employees in each school (P1: 45, P2: 75, P3: 56, P4: 59, C5: 37, and C6: 40 employees) fulfilled the applicable criteria of minimum of 20 full-time employees involved in the survey (FIOH 2020). The average response rate was 68% (P1: 85%; P2: 69%; P3: 70%; P4: 73%; C5: 46%; and C6: 65%), being at the same level than the response rate (average about 70%; range 40–100%) in other studies utilizing the FIOH's questionnaire (Tähtinen et al. 2019, Reijula and Sundman-Digert 2004, Pasanen et al. 2002). Although the moderately low response rate may weaken the reliability of the results in school C5 and the possibility of self-selection bias cannot be eliminated, results can still be considered important indicative signals. Higher sensation of symptoms and poor IAQ could motivate the occupants to respond to questionnaires in problematic schools. In a study regarding family-administered questionnaires, the families having greater number of reported symptoms from the child and other family members were more inclined to participate in the study (Bornehag et al. 2006).

The perceived work environment and symptoms in each school based on the FIOH's indoor climate survey are presented in Fig. 2. More detailed results are available in Table S2 in the supplementary material.

As shown in Fig. 2 and Table S2, the work-related symptoms were clearly more prevalent in the problematic schools than in the comparison schools. Less than 20% of responses from the comparison schools reported symptoms, complying with the reference values, whereas in the problematic schools, symptom prevalence was higher. In school C5, occupants reported more tiredness (19%), hoarse or dry throat (18%) and skin symptoms of hands (18%) compared to reference values. However, altogether in the problematic schools, higher number of occupants reported these symptoms (16–37%, 28–46% and 13–26%, respectively). It has been shown that the subjective IAQ is more based on building-related factors in schools with indoor air problems, and in schools without IAQ problems it may reflect psychosocial factors (Finell et al. 2017). In school C5, occupants reported a lot of stress, too much work and poor possibilities to effect own work.

Work environment-related discomfort was also reported less in comparison schools; however, in general, the difference between problematic and comparison schools was not so clear than in symptoms. In school C5, occupants reported stuffy air and insufficient ventilation which might be explained by the low air flow rates (see Section 3.2). In school C6, high temperature was reported more than in other schools. Higher temperatures in this school were supported also by the temperature measurements (see Section 3.2). In school P2, occupants reported only temperature-related discomfort and otherwise seemed to be satisfied with the work environment compared to the reference values. Noise and poor lighting or reflections are not indoor air-related factors as such but are affecting the holistic perception of personal wellbeing in the workplace; therefore, the results are presented here alongside the other results. Noise was reported in all schools, and poor lighting or reflections only in school C5.

Based on the occupants' perceptions of both work environment and symptoms, schools P1 and P3 seemed to have most IAQ problems; however, the symptoms in P2 and P4 were also clearly elevated when compared to both the reference data and comparison schools. The relative humidity of indoor air was very low in all the schools, particularly in school P3 (see Section 3.2), which could have affected the perceived symptoms and discomfort (Wolkoff 2018, Wolkoff and Kjærgaard 2007). Perceptions related to stuffy or dry air and insufficient ventilation, and symptoms related to eye or nose irritation, hoarse throat or cough, difficulties in breathing or facial or hand skin

symptoms were emphasized in problematic schools. These are all factors that could be related to the cold season and dry indoor air. Furthermore, feeling heavy-headed, tiredness and headaches were prevalent especially in problematic schools. Increased reporting of especially respiratory (i.e., runny nose, stuffy nose, cough, hoarseness, and sore throat) and general symptoms (i.e., fatigue, concentration difficulties, and headache) have been found to correlate with expert-assessed indoor environment quality problems in a previous study (Savelieva et al. 2019). In addition, in schools P1 and P2, self-reported stress, other psychosocial factors and asthma were more prevalent than in the reference data (Reijula and Sundman-Digert 2004), possibly increasing symptom perceptions.

Occupants' worry about perceived IAQ problems had been present in all problematic schools over a longer period, negatively affecting the atmosphere at work. This kind of work environment-related psychosocial stress is known to cause more complaints and symptoms regarding the indoor environment (Lahtinen et al. 2004, Brauer and Mikkelsen 2010), and the role of worry and socioemotional difficulties in increased symptom perceptions have been shown recently in several studies (Nissilä et al. 2019, Finell et al. 2018). Also, knowledge of an ongoing study may considerably affect the reporting of symptoms (Brauer and Mikkelsen 2003). At the beginning of this study, the occupants were told whether their school was considered a problematic or a comparison school. Therefore, it is possible that occupants who had more symptoms were overrepresented in this study.

3.1.2. Online questionnaire for the students and teachers

The online questionnaire results were detailed and compared to measurements of temperature, relative humidity and CO₂ in Järvi et al. (2018). In the problematic schools, teachers and students reported more symptoms and inconveniences than in comparison schools. Differences were significant for symptoms and discomfort reported by teachers and for symptoms reported by students during the last hour (*p*-values < 0.001). The results indicated that the most problems occurred in schools P3 and P4. Many significant differences occurred between the classrooms in both problematic and comparison schools (Järvi et al. 2018).

The results of the online questionnaire confirm those of the FIOH's questionnaire, with some exceptions. It must be noted that these two questionnaires gave a different perspective to the studied buildings: the online questionnaire provided information only from the studied classrooms during a specific timeframe, whereas the FIOH's questionnaire covered the whole building and addressed long-time perceptions. Moreover, the respondent groups were different, with only a group of teachers and students completing the online questionnaire and employees throughout the building completing the FIOH's questionnaire.

3.2. Ventilation and physical parameters

Differential pressures and air flow rates are shown in Table S3 in the supplementary material. The sufficiency of the supply air rate was calculated by dividing the supply air flow rate by the required design air flow rate 6 dm³/(s person) (Ministry of the Environment, Finland 2017). Air flow rates could be measured during the visits to schools P2, P3, P4 and C5, excluding classrooms 2b, 4f, 5a, 5b and 5f. The supply air flow rates were considered sufficient in most of these classrooms (sufficient for 21–48 people), except in C5, where the ventilation was sufficient for only 11–15 people, and P4, where the ventilation was sufficient for only 15–27 people. These occupancy rates are likely exceeded on occasion, since the typical number of students per class is 20–30 people.

In the majority of the classrooms (66%), dPa values were moderately negative (max. −10 Pa) from −9.3 Pa to −0.7 Pa. Negative dPa from −10.2 Pa to −18.7 Pa occurred in 21% of the classrooms (1b–d, 3f, 4c–e and 5e), and moderate positive dPa from 3.5 Pa to 8.0 Pa

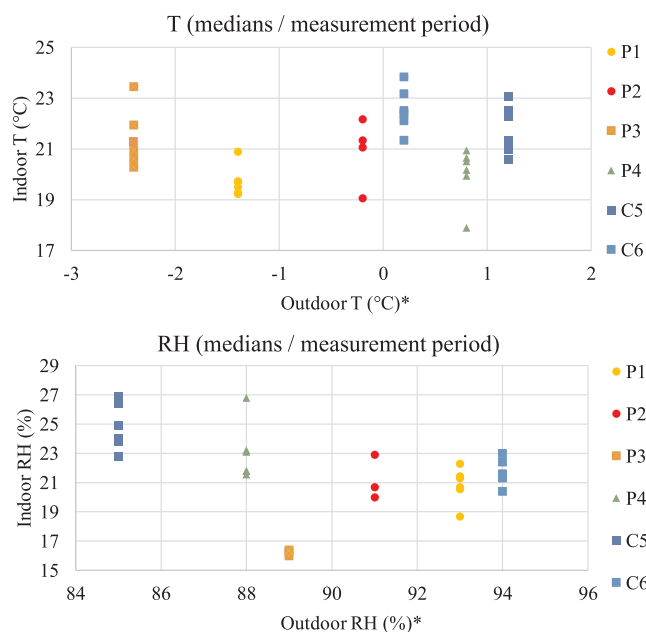


Fig. 3. Temperature and relative humidity in the classrooms presented as medians and as a function of median outdoor temperature and relative humidity during the measurement periods. *Finnish Meteorological Institute (2019).

occurred in 13% of the classrooms (2a, 2b, 2e, 3e and 6f). Ergo, high negative dPa values (over -10 Pa) were established mostly in schools P1 and P4, and positive dPa established mostly in school P2. According to the national decree, high negative dPa should be avoided to prevent the infiltration of possible harmful impurities inside building structures (Ministry of the Environment, Finland 2017). Positive differential pressure might prevent the infiltration (Vornanen-Winqvist et al. 2018a).

T, RH and CO_2 in the classrooms during school days, between the working hours of 8 am and 5 pm (median, min and max values) are shown in Table S4 in the supplementary material. Medians of T and RH during the two weeks measurement period are shown as a function of median outdoor temperature and relative humidity in Fig. 3.

From 8 am to 5 pm, schools P3, C5 and C6 had the highest median temperatures (21.9, 22.6 and 21.9 °C, respectively), and school P1 had the lowest one (20.4 °C). However, all the schools' indoor temperatures were within the limits of $+20$ °C (min) and $+26$ °C (max) during the heating season (Ministry of Social Affairs and Health, Finland 2015).

Low RH values were typical for the cold season, especially in well-ventilated buildings. RH was lowest in the P3 classrooms (median between 8 am and 5 pm: 16.1%) and highest in the C5 classrooms (median: 24.5%). In other schools, median RH values were between 21% and 23%. RH of about 40% is suggested for the well-being of eyes and upper airways (Wolkoff and Kjærgaard 2007). In particular, excessively dry indoor with RH below 20% for a long time is known to cause irritation and symptoms (Fang et al. 2004, Sunwoo et al. 2006), and a low relative humidity of 20–35% also favours the spread of influenza viruses (Lowen et al. 2007, Kudo et al. 2019). Therefore, the low RH values could contribute to some symptoms observed by the occupants, especially in school P3.

From 8 to 5 pm, schools C6 and P1 were characterised by higher median CO_2 concentrations (564 and 520 ppm, respectively) compared to schools P2, P3, P4 and C5 (458, 466, 443 and 482 ppm, respectively). In comparison schools, more classrooms reached high maximum CO_2 concentrations than in problematic schools. The limit for indoor CO_2 is 1,150 ppm higher than outdoors (approx. 400 ppm) (Ministry of Social Affairs and Health, Finland 2015), which was not exceeded in any of the classrooms.

3.3. Gaseous pollutants and particulate matter

VOC and TVOC concentrations detected in air samples are presented in Table S5 in the supplementary material. TVOC concentrations were $< 10\text{--}30$ $\mu\text{g}/\text{m}^3$ (school P1), $< 10\text{--}20$ $\mu\text{g}/\text{m}^3$ (P2), $< 10\text{--}40$ $\mu\text{g}/\text{m}^3$ (P3), $< 10\text{--}20$ $\mu\text{g}/\text{m}^3$ (P4), $20\text{--}30$ $\mu\text{g}/\text{m}^3$ (C5) and $30\text{--}50$ $\mu\text{g}/\text{m}^3$ (C6). These concentrations were higher in the comparison schools (averages 22 (C5) and 37 (C6) $\mu\text{g}/\text{m}^3$) than in the problematic schools (12 (P1), 9 (P2), 8 (P3) and 12 (P4) $\mu\text{g}/\text{m}^3$), and interestingly, a greater variety of VOCs were detected in the comparison schools. In all studied classrooms, TVOC and single VOC concentrations were below the national decree's reference values (400 $\mu\text{g}/\text{m}^3$ for TVOC and $10\text{--}50$ $\mu\text{g}/\text{m}^3$ for single compounds), except for classroom 3b, where the ethanol concentration (75 $\mu\text{g}/\text{m}^3$) exceeded the action limit of 50 $\mu\text{g}/\text{m}^3$ (Ministry of Social Affairs and Health, Finland 2015).

The major compound groups detected were aromatic hydrocarbons, alcohols and aldehydes. The median concentration of 69% of all detected compounds was ≤ 1 $\mu\text{g}/\text{m}^3$. Benzene (median concentration 0.5 $\mu\text{g}/\text{m}^3$), toluene (0.5 $\mu\text{g}/\text{m}^3$), nonanal (1 $\mu\text{g}/\text{m}^3$), 2-ethyl-1-hexanol (2 $\mu\text{g}/\text{m}^3$), benzaldehyde (1 $\mu\text{g}/\text{m}^3$), decanal (1 $\mu\text{g}/\text{m}^3$), 1-butanol (1 $\mu\text{g}/\text{m}^3$), *p,m*-xylenes (0.5 $\mu\text{g}/\text{m}^3$), decamethylcyclotrisiloxane (2 $\mu\text{g}/\text{m}^3$) and acetone (3 $\mu\text{g}/\text{m}^3$) were the most commonly found compounds in the classrooms.

The measured TVOC concentrations in the present study were similar to those observed in 144 mechanically ventilated classrooms in the USA, with concentrations of $10.12\text{--}45.89$ $\mu\text{g}/\text{m}^3$ during winter measurements (Zhong et al. 2017), and lower than 73 naturally ventilated Portuguese classrooms with a TVOC median of 140.3 $\mu\text{g}/\text{m}^3$ (Madureira et al. 2009) or a review study of 104 classrooms with 60% of the values between 80 and 200 $\mu\text{g}/\text{m}^3$ (Chatzidiakou et al. 2012).

Benzene, toluene and *p,m*-xylene were also among the most observed VOCs in 144 American classrooms, with median concentrations of 0.3, 0.7 and 0.4 $\mu\text{g}/\text{m}^3$, respectively (Zhong et al. 2017). In that study, aromatics, alkanes and terpenes were the major compound groups detected, and most VOCs' mean concentrations were below 5 $\mu\text{g}/\text{m}^3$ (Zhong et al. 2017), similar to our results. In the 73 Portuguese classrooms, meanwhile, the median concentrations were 2.5 $\mu\text{g}/\text{m}^3$ for benzene and 6.4 $\mu\text{g}/\text{m}^3$ for toluene (Madureira et al. 2015).

VOC emissions in schools originate from a combination of indoor building materials, human activities and outdoor sources (Chatzidiakou et al. 2012). Cleaning products and art and craft activities are typical sources for elevated VOC concentrations in schools (Mishra et al. 2015, Śmiełowska et al. 2017). Art activities took place in classrooms 1d, 2c and 3f; thus, the teachers and students in these classrooms were more exposed to VOC-containing teaching materials than in the other classrooms. In classroom 2c, 2-ethyl-1-hexanol (4 $\mu\text{g}/\text{m}^3$) and nonanal (2 $\mu\text{g}/\text{m}^3$) were detected. Art classroom 3f had higher TVOC concentration (40 $\mu\text{g}/\text{m}^3$) compared to the other P3 classrooms (≤ 10 $\mu\text{g}/\text{m}^3$), consisting of numerous compounds with concentrations of < 2 $\mu\text{g}/\text{m}^3$. Classroom 1d showed no difference to the other classrooms in the same school.

Toluene in classrooms 4d and 4e; 2-phenoxyethanol in classrooms 2c, 5a, 5b, 5c and 5e; and propylene glycol in classroom 2c exceeded the recommended limit values for workplaces published by the FIOH (2017). Repeated VOC-samplings would be necessary to confirm their long-term presence in the visited classrooms.

Formaldehyde concentrations were under the Finnish action limit of 50 $\mu\text{g}/\text{m}^3$ (Ministry of Social Affairs and Health, Finland 2015). Specifically, measurements were below the device's detection limit of 10 ppb (equal to approximately 12 $\mu\text{g}/\text{m}^3$) in schools C6 and P4 and reached a maximum of 13 ppb (equal to approximately 16 $\mu\text{g}/\text{m}^3$) in the other schools. These results were in the range of formaldehyde concentrations found in past school studies. For example, a median formaldehyde concentration of 6.0 $\mu\text{g}/\text{m}^3$ had been reported in mechanically ventilated American school buildings (Zhong et al. 2017) and a 17.5 $\mu\text{g}/\text{m}^3$ concentration had been reported in naturally

Table 2Average, median, minimum and maximum values of PM_{2.5} and PM₁₀ concentrations in the studied classrooms on school days between 8 am and 5 pm.

School	PM _{2.5} [µg/m ³]						PM ₁₀ [µg/m ³]					
	P1	P2	P3	P4	C5	C6	P1	P2	P3	P4	C5	C6
Classroom	1e	2g	3a	ND	5b	ND	1e	2g	ND	ND	5b	6d
Total measurement time [days]	14	16	4		14		2	16			14	3
Average	2.7	1.0	2.3		2.4		3.4	2.5			3.1	1.8
Median	2.1	0.5	2.1		2.2		3.3	1.8			2.8	1.5
Min	0.5	0.0	0.7		0.3		2.4	0.4			0.6	0.8
Max	24.2	24.9	4.3		6.2		6.8	24.1			11.2	11.5

ND: no data.

ventilated Portuguese buildings (Madureira et al. 2015).

Neither ammonia, arsine, ozone nor hydrogen sulphide were found in the indoor air of the studied schools. CO varied from 0.1 to 1.3 ppm, which is fairly below the Finnish action limit of 7 mg/m³ (equivalent to 6.0 ppm) (Ministry of Social Affairs and Health, Finland 2015) and WHO (2010) guidelines of 6.1 ppm for a 24-hour exposure. Research on CO levels in school classrooms is limited, but some epidemiological studies have associated mean CO levels of 2.73 ppm (range: 0.65–6.23 ppm) with an approximately 4.0% increase in absenteeism with no lagged effects (Chen et al. 2000, Currie et al. 2009). CO is a colourless, odourless toxic gas formed by incomplete combustion of organic matter, and in most of the school studies, the concentrations were low, since CO mainly originates from outdoor vehicular emissions (Chithra and Nagendra 2018).

Fine and coarse particulate matter (PM_{2.5} and PM₁₀) concentrations are shown in Table 2.

PM concentrations were low compared to the action limits of 25 µg/m³ for PM_{2.5} and 50 µg/m³ for PM₁₀ (Ministry of Social Affairs and Health, Finland 2015). Daily recommended WHO guidelines are 20 µg/m³ for PM₁₀ and 10 µg/m³ for PM_{2.5} (WHO 2010). The medians varied from 0.5 to 2.2 µg/m³ (PM_{2.5}) and from 1.8 to 2.8 µg/m³ (PM₁₀) during school hours. Curiously, in school C5, the median concentrations were the highest among the studied schools. However, multiple peaks of PM_{2.5} and PM₁₀ concentrations occurred in schools P1 and P2.

The results of the present study are lower than those found in European classrooms (mean daily PM_{2.5} concentration of 20.2 µg/m³ and PM₁₀ concentrations from 43 to 169 µg/m³) (Chatzidiakou et al. 2012) and school studies in different climatic regions (median of 24 µg/m³ for PM_{2.5} and of 105 µg/m³ for PM₁₀) (Salthammer et al. 2016). The studied schools' cleanliness, ambient environment and proper ventilation likely promoted their low PM concentrations.

3.4. Endotoxin loads in settled surface floor dust

Endotoxin concentrations in each visited classroom are presented in Table 3.

Compared to the concentrations reported previously in literature, the endotoxin levels of the present study were extremely low. The average endotoxin concentrations in schools P1, P2, C5 and C6 (1.5, 7.3, 1.1 and 15.8 EU/m², respectively) were relatively low compared to schools P3 and P4 (77.5 and 56.4 EU/m², respectively). According to Salonen et al. (2016), endotoxin levels in school settled dust may vary between 2180 and 48,000 EU/m². In the HITEA study, mean concentrations of 1100–3300 EU/m² were reported from the Finnish (Jacobs et al. 2014a,b). However, in the HITEA study, the dust was collected for eight weeks which is a major difference compared to this study.

Endotoxin levels are affected by several factors, such as the building's age, cleaning, flooring materials, ventilation system, presence of animals, relative humidity and temperature (Salonen et al. 2013, Jacobs et al. 2014a,b). These factors contributed to low endotoxin concentrations in the present study: the schools had plastic flooring and

Table 3Loads of endotoxin (EU/m²) in the settled floor dust of the six studied schools.

Endotoxins [EU/m ²]	P1	P2	P3	P4	C5	C6
<i>Classroom</i>						
a	–	4.9	89.3	8.3	0.9	3.5
b	3.7	1.1	192.3	132.6	0.5	9.3
c	0.5	1.7	9.9	15.3	0.5	16.1
d	–	1.0	36.9	67.4	3.1	7.6
e	1.4	7.4	76.9	23.2	0.9	46.8
f	0.6	6.1	59.6	91.4	1.0	11.4
g		13.2				
h		23.0				
Field blank	1.3	2.7	3.9	3.2	–	2.9
Average	1.5	7.3	77.5	56.4	1.1	15.8
Median	1.0	5.5	68.2	45.3	0.9	10.3
Std	1.5	7.5	63.0	49.5	1.0	15.8
Min	0.5	1.0	9.9	8.3	0.5	3.5
Max	3.7	23.0	192.3	132.6	3.1	46.8

other hard-surfaced indoor materials; there were no animals, and despite the few days long cleaning break, regular cleaning activities were in place. In addition, the buildings were relatively new or renovated, the RH of indoor air was low (medians from 15% to 27%), outdoor temperatures were low and indoor temperatures were rather uniform between schools.

The data variability between classrooms was quite high, especially in schools P3 and P4 (standard deviations of 63.0 and 49.5, respectively). In schools C5 and P1, variability between classrooms was low (1.0 and 1.5, respectively). Classroom use and occupancy following the latest cleaning could explain the differences and data variability. For example, in classrooms 2c and 2e, normal cleaning activities were conducted without the requested cleaning break, and endotoxin levels were lower compared to the other rooms. In contrast, the endotoxin levels were high in classrooms 3a and 3b, in which mice excrement was observed daily, according to the teacher, and during the measurement visit.

Single endotoxin measurements may be used to estimate annual endotoxin levels in schools (Jacobs et al. 2014a,b). However, previous research data indicate that although endotoxin quantification in floor settled dust is generally accepted, the measurements of endotoxins in air samples might better represent inhaled endotoxins (Noss et al. 2008, Park et al. 2006, Jacobs et al. 2014a,b).

3.5. Airborne cultivable microorganisms

Concentrations of airborne cultivable microorganisms and indoor/outdoor (I/O) ratios in the studied problematic and comparison schools are shown in Table 4.

Differences between classrooms were especially notable in schools P2–P4. P1 was left out of the statistical evaluation since the outdoor microbial concentration was not available.

An I/O ratio > 1 might indicate indoor sources of microbes in buildings, and if I/O ratio is close to 1, the microbial concentrations are

Table 4

Concentrations of airborne cultivable microorganisms and indoor/outdoor (I/O) ratios in the studied schools P1–C6.

Concentrations [cfu/m ³] Classroom	P1 TC	I/O	P2 TC	I/O	P3 TC	I/O	P4 TC	I/O	C5 TC*	I/O	C6 TC	I/O
a	170	–	10	0.1	250	5.0	160	0.5	0	0.0	140	1.2
b	240	–	–	–	190	3.8	90	0.3	30	0.2	100	0.8
c	240	–	50	0.5	540	10.8	260	0.8	20	0.1	60	0.5
d	250	–	510	5.1	360	7.2	100	0.3	10	0.1	30	0.3
e	160	–	0	0.0	180	3.6	540	1.7	60	0.4	90	0.8
f	1680	–	290	2.9	40	0.8	410	1.3	10	0.1	100	0.8
g			370	3.7								
h			230	2.3								
Average	212		209		260		260		22		87	
Median	240		230		220		210		15		95	
Std	43		197		172		182		21		38	
Outdoor	–		100		50		320		170		120	
	YM	I/O	YM	I/O	YM	I/O	YM	I/O	YM	I/O	YM	I/O
a	10	–	0	0.0	30	1.0	10	0.1	0	0.0	0	–
b	0	–	–	–	0	0.0	40	0.3	0	0.0	0	–
c	100	–	20	0.1	10	0.3	10	0.1	0	0.0	0	–
d	10	–	10	0.0	0	0.0	0	0.0	0	0.0	0	–
e	0	–	10	0.0	0	0.0	10	0.1	20	0.1	0	–
f	0	–	0	0.0	0	0.0	30	0.2	10	0.1	0	–
g			20	0.1								
h			0	0.0								
Average	20		9		7		17		5		0	
Median	5		10		0		10		0		0	
Std	39		9		12		15		8		0	
Outdoor	–		390		30		160		180		0	

TC: total cultivable microorganisms, YM: yeast and mould. Classroom 1f (TC) concentration was not included in statistical calculations. *TC after 7 days.

generally similar in indoor and outdoor air (Salonen et al. 2015). The I/O ratio (total count) was > 1 in most P2 and P3 classrooms, in two P4 classrooms, and in one C6 classroom. The I/O ratio (yeast and mould) was < 1 in all schools, except one classroom in P3.

Elevated total count concentrations in problematic schools may indicate microorganism sources indoors and consequently require additional environmental investigations. Concentrations could also be elevated due to occupants, particularly bacteria, if the classrooms had poor ventilation; however, the air flow rate measurements (see Section 3.2) did not support this suggestion.

Climatic conditions are the determinants of airborne viable fungal concentrations (Salonen et al. 2015, Pyrrri and Kapsanaki-Gotsi 2017). In the subarctic region, differences between damaged and non-damaged buildings are more obvious during winter than other seasons with higher outdoor fungal concentrations (Meklin 2002). The space–time variation of airborne viable fungal concentrations in the outdoor environment is high (WHO 2009), and data interpretation should be based on multiple air sampling campaigns (Salonen et al. 2015). Therefore, repeated samples and longitudinal studies are necessary to confirm airborne microbial exposure in an occupational setting.

3.6. Characterisation of the cultivable moulds in indoor airborne and settled dust with special focus on toxigenic *Trichoderma* strains

The categorisation of indoor fungal colonies into morphotypes and identification to the species level of morphotypes representing the genera *Trichoderma*, *Rhizopus*, *Chaetomium*, *Dichotomophilus* and *Fusarium* are summarised in Table S6 in the supplementary material. *Aspergillus*, *Eurotium*, *Paecilomyces* and *Penicillium* morphotypes are described in Vornanen-Winqvist et al. (2018a, 2018b).

The dominance and occurrence of each fungal species, genus or morphotype in each school were estimated based on the frequency of plates recorded positive for growth, since overgrowth of fungal genera, such as *Rhizopus*, prevented colony quantification. Frequently occurring fungal colonies of toxigenic morphotypes belonged to the genera

Trichoderma, *Aspergillus*, *Penicillium*, *Fusarium* and *Chaetomium*. Potentially pathogenic strains able to grow at 37 °C were identified as *Trichoderma* (*T. citrinoviride*, *T. longibrachiatum*), *Aspergillus* (*A. flavus*, *Aspergillus* sp. section *Nigri*), *Rhizopus* (*R. oryzae*) and *Paecilomyces* sp. Non-toxicogenic and non-pathogenic *Rhizopus* and *Penicillium* strains were also present in the indoor dust in all six studied schools. The results are presented in Table 5.

Table 5 shows that toxigenic *Trichoderma* strains were found on 38% of the 52 plates inoculated with airborne and settled dust collected from schools P1–P4, but not on any of the 21 plates associated with schools C5 and C6, indicating an incidence of < 5%. The problematic schools associated with IAQ complaints shared toxigenic representatives of the genus *Trichoderma*: the mycoparasitic species *T. atroviride* (Druzhinina et al. 2011, Seidl et al. 2009) and *T. trixiae* and the potentially pathogenic species *T. longibrachiatum* and *T. citrinoviride*. *T. trixiae* has been documented once in a Finnish indoor environment, while *T. atroviride* was the most frequent *Trichoderma* species isolated from Finnish buildings (Castagnoli et al. 2018).

Mycoparasitically and mycotrophically growing *Trichoderma* produce many biologically active substances participating in the mycoparasitic process, as proteases, chitinases and peptaibols (Druzhinina et al. 2011). Trichorzianins and trilogins, peptaibols from *Trichoderma* toxic to mammalian cells in *in vitro* and *ex vivo* assays, are emitted as liquid-containing vesicles by indoor *T. atroviride* and *T. trixiae* strains (Castagnoli et al. 2018). Biologically active metabolites could therefore be airborne in buildings showing *Trichoderma* growth.

Toxigenic and potentially pathogenic *Aspergillus* strains were found in schools P3, P4, C5 and C6. Toxigenic *Chaetomium globosum* and non-toxicogenic *Dichotomophilus* sp. strains were detected in school C5, while toxigenic *Fusarium* sp. strains were the dominant colonisers of plates from school C6. Our results indicate that viable conidia of toxigenic and potentially pathogenic mould species may also be found in healthy buildings or buildings where health symptoms have not evolved.

A summary of the found toxigenic and/or pathogenic moulds in the six studied schools is presented in Table 6. Tables 5 and 6 show that

Table 5

Toxigenic *Trichoderma* strains and other frequent fungal species isolated on MEA from exhaust filters of air handling units and settled dusts collected in six Finnish representative schools. Toxigenic strains were screened toxic in *ex vivo* and /or in vitro assays.

Sampling		Toxigenic <i>Trichoderma</i> species		Other frequent fungal species and genera		
Time and material		Identified strains (plates positive/all plates)		Identified toxigenic strains (plates positive/all plates)		Other identified strains (plates positive/all plates)
P1	Nov 2016	EF	<i>T. atroviride</i>	6/6		<i>Rhizopus oryzae</i> ¹ 1/6 <i>Rhizopus</i> sp. 5/6 <i>Penicillium</i> sp. 3/6
P2	Nov 2016	EF	<i>T. trixiae</i>	6/9		<i>Rhizopus</i> sp. 5/9 <i>Penicillium</i> sp. 5/9
P3	Aug 2016	EF	<i>T. atroviride</i>	2/8	<i>Pen. chrysogenum</i> 7/8 <i>Pen. commune</i> 7/8 <i>Asp. section Nigri</i> ¹ 7/8 <i>Paecilomyces</i> sp. ¹ 1/8 <i>Asp. westerdijkiae</i> 3/9 <i>Pen. expansum</i> 1/9 <i>Asp. versicolor</i> 2/2	<i>Rhizopus</i> sp. 3/8 <i>Asp. flavus</i> ¹ 2/8 <i>Penicillium</i> sp. 1/8
	Feb 2017	EF ³	<i>T. longibrachiatum</i> ¹	1/9		<i>Rhizopus</i> sp. 4/9 <i>Penicillium</i> sp. 6/9
	May 2016	SD	<i>T. atroviride</i>	1/2		<i>Rhizopus</i> sp. 1/2
	Sep 2016	SD ³	<i>T. trixiae</i>	1/3		<i>Rhizopus</i> sp. 2/3 <i>Penicillium</i> sp. 2/3
P4	March 2017	EF	None	< 1/6	<i>Asp. westerdijkiae</i> 1/6	<i>Penicillium</i> sp. 6/6
	May 2016	SD ⁴	<i>T. citrinoviride</i> ¹	1/6		<i>Rhizopus</i> sp. 2/6
			<i>T. atroviride</i>	2/6		
	March 2017	SD	None	< 1/9	<i>Asp. westerdijkiae</i> 2/9 <i>Asp. section Nigri</i> ¹ 3/9	<i>Penicillium</i> sp. 3/9 <i>Penicillium</i> sp. 2/9 <i>Eurotium</i> sp. 1/9
C5	Nov 2016	EF	None	< 1/9	<i>Chaetomium globosum</i> ² 3/9 <i>Asp. section Nigri</i> ¹ 2/9	<i>Rhizopus</i> sp. 2/9 <i>Dichotomophilus</i> sp. 6/9 <i>Penicillium</i> sp. 6/9
C6	Jan 2017	EF	None	< 1/6	<i>Fusarium</i> sp. 6/6 <i>Asp. westerdijkiae</i> 2/6 <i>Asp. section Nigri</i> ¹ 1/6	<i>Rhizopus</i> sp. 1/6 <i>Penicillium</i> sp. 2/6
		SD	None	< 1/6	<i>Fusarium</i> sp. 4/6 <i>Asp. westerdijkiae</i> 1/6	<i>Rhizopus</i> sp. 1/6 <i>Penicillium</i> sp. 1/6

EF: exhaust filter of the air handling unit; SD: settled dust from the classroom surfaces

¹ Able to grow at 37 °C.

² Unable to grow at 37 °C but reported to infect immunocompromised patients (Lesire et al. 1999).

³ During September 2016 and February 2017, a continuous positive differential pressure of about 5 Pa was generated in the building part served by the investigated air handling unit.

⁴ Before sampling, the supply air income had been insufficient for several months, due to a control system error.

potentially pathogenic and toxigenic fungi were detected in every school, but the diversity of the toxigenic species varied between the schools.

In Finland, boron chemicals and cationic polyguanide type of biocidal disinfectants, such as PHMB, are used in construction materials to remediate mould-affected indoor spaces and prevent future mould growth in new buildings (Mikkola et al. 2015). Strains representing different species of the genus *Trichoderma* and selected reference strains were therefore tested for resistance against the biocides PHMB and borax. Three resistance patterns were found and are shown in Fig. 4. A summary of the species diversity of the strains belonging to the genus *Trichoderma* and their resistance patterns towards borax and PHMB are shown in Table S6 in the supplementary material.

The resistance towards PHMB and borax of *T. trixiae* strains was visible after one week of incubation. *T. atroviride* strains also expressed resistance towards borax following one week but needed three weeks to fully grow and sporulate on PHMB plates. The growth of *T. citrinoviride* and *T. longibrachiatum* strains was inhibited by PHMB and borax, even after three weeks of incubation. Interestingly, the outdoor reference strains of *T. atroviride* (SZMC 1723 and SZMC 12495) were also resistant to borax and PHMB. The resistance towards borax and PHMB recorded in *T. atroviride* seemed to be a species-specific trait, rendering the species well-adapted to grow in indoor environments contaminated with these biocides.

Results in Tables 5 and 6 and Fig. 4 show that borax- and PHMB-resistant strains of the species *T. trixiae* and *T. atroviride* were detected

Table 6

Summary of the found toxigenic and/or pathogenic moulds in the six studied schools.

School	<i>Trichoderma</i> strains growing at		Toxigenic and/or pathogenic <i>Aspergillus</i> , <i>Paecilomyces</i> and <i>Penicillium</i> strains	<i>Rhizopus</i> sp. growing at		Toxigenic strains of	
	22 °C, toxigenic	37 °C		37 °C	22 °C	<i>Fusarium</i> sp.	<i>Chaetomium</i> sp.
P1	+	–	–	+	+	–	–
P2	+	–	–	–	+	–	–
P3	+	+	+	–	+	–	–
P4	+	+	+	–	+	–	–
C5	–	–	+	–	+	–	+
C6	–	–	+	–	+	+	–

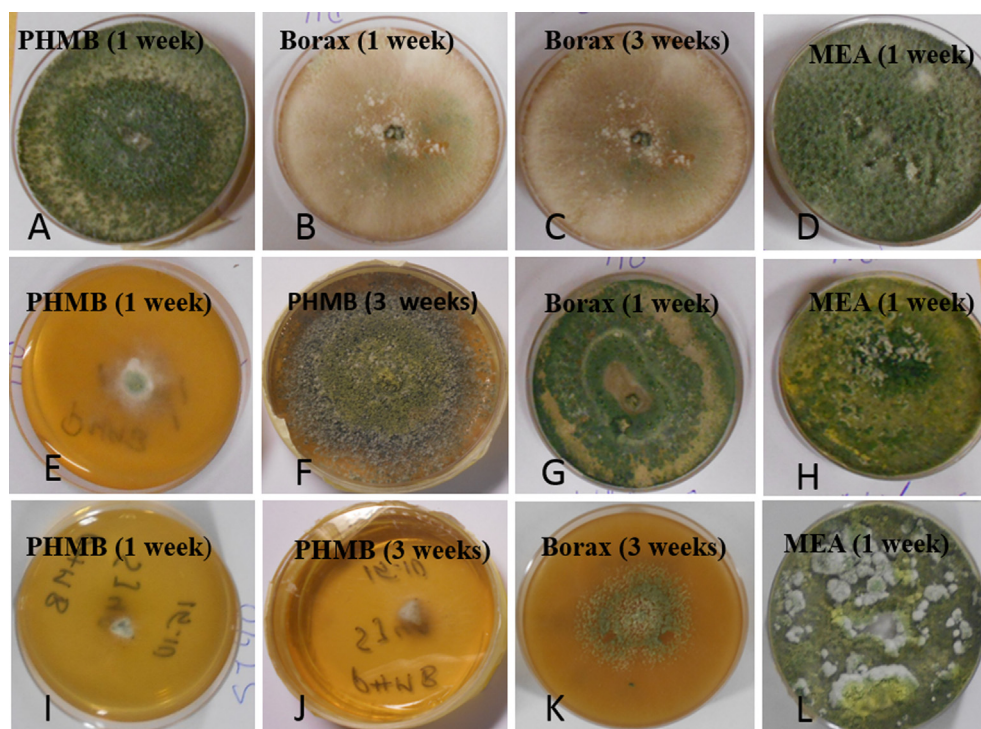


Fig. 4. Resistance patterns to PHMB and borax of indoor *Trichoderma* isolates. Strains were considered resistant (R) if able to completely cover the plate with sporulating hyphae and sensitive (S) if unable to cover the plates within three weeks. The rows show the resistance patterns exhibited by the *Trichoderma trixiae* strains (A, B and C), the *T. atroviride* strains (E, F and G) and the potentially pathogenic *T. citrinoviride* and *T. longibrachiatum* strains (I, J and K). Non-inoculated MEA plates (not shown) and growth on non-supplemented biocide MEA plates (D, H and L) were used as controls.

in schools P1–P4, whereas the biocide-sensitive strains of potentially pathogenic *T. longibrachiatum* and *T. citrinoviride* were found in P1 and P2, respectively.

The mycoparasitism and mycotrophy of the genus *Trichoderma*, and especially species *T. atroviride*, indicate adaptation to environments contaminated with other fungi (Druzhinina et al. 2011, Seidl et al. 2009). Thus, the presence of cultivable conidia of toxigenic, biocide-tolerant and mycoparasitic *Trichoderma* species in a building may indicate long-term excess moisture and mould growth of other fungi, as well as emission of toxic and immunoreactive metabolites into the indoor air. This study refers that moisture and mould problems were possibly present in the problematic schools and might have correlated with the perceived symptoms and discomfort. However, the moisture and mould problems might have been evidently hidden and difficult to investigate, since severe problems were not found from the schools in the building investigations during the past years.

It must be noted that the culture of dust collected from exhaust air filters does not provide information about the origin of the spores or location of the possible growth. For example, the presence of outdoor fungi in the filter dust is possible, due to windows kept open in the classrooms; however, the absence of the most common outdoor fungi *Cladosporium* sp. (Salonen et al. 2015, Barnes 2019) indicated that the filter dust was a good representative of the indoor air. Teachers and students' contingent exposure to pathogenic and/or toxigenic moulds needs to be confirmed with proper sampling and structural investigations. Simple and effective long-term air and surface samplings are needed for proper building investigations. In addition, ventilation-related parameters, such as dPa and low relative humidity of indoor air, could have a notable effect on the occupants' wellbeing and therefore should be addressed more precisely in the future studies.

3.7. Advantages and limitations of the study

To the authors' knowledge, this is the first study to investigate school buildings with and without indoor air problems using the combined analysis of gaseous, particulate and biological pollutants, physical parameters and the occupants' perceptions recorded by two different types of questionnaires. The study applied several punctual and

long-term measurements, including commonly used methods and less common approaches such as detecting microbes from exhaust air filters and implementing a new, fast online questionnaire.

Since the FIOH's Örebro (MM40) -based indoor climate survey is commonly used, responses collected in the present study were compared to a well-established reference database from studying Finnish workplaces. However, analysis was limited since office environments have some obvious differences compared to schools, for example, in occupancy, activities and noise. Furthermore, the limited number of respondents decreased the reliability of the data, but an acceptable average response rate was reached. Regarding the online questionnaire, several responses from the same respondents were collected. In addition, the questionnaire was short and provided data about up-to-date perceptions and exact locations of possible sources for the perceived problems. By implementing the questionnaire to all occupants of the school, all rooms could be included in the study, which would provide undisputed benefits as a support material for building investigations.

A different perspective to the studied buildings was given by the two questionnaires. The indoor climate survey collected information on long-time perceptions from the employees of the building, whereas the online questionnaire provided information about children's and teachers' perceptions on classrooms for a specific period. According to the authors, for recording the occupants' perceptions, an online questionnaire collecting up-to-date data continuously from different rooms and different groups, such as students and teachers, seems more beneficial than traditional questionnaires.

The number of samples was limited in this study, but many represented long-term measurements. Such measurements were, for example, the exhaust air filters and settled dust and endotoxin samples, as well as the two-week monitoring of particulate matter, formaldehyde, and CO₂. Sampling methods were chosen in order to collect as many pollutants as possible in the short period available for the conducting the study.

In the present study we collected both settled dust samples as well as air samples from indoor and outdoor air. Based on the microbial analyses, differences between the problematic and comparison schools were found, showing an indication but not causative relationships between the reported health problems and indoor microbes.

The active sampling of airborne dust is expected to be the most representative way to evaluate the microbial exposure for building occupants. However, the sampling of settled dust is also widely used as a surrogate of air sampling in studies assessing the health effects of indoor microbes (Leppänen et al. 2018). The dust sampling generally represents a long-term enrichment of settled airborne microbes. The amount of airborne dust in these schools has not been measured, and, furthermore, the actual concentrations and duration of the exposure to the settled dust associated microbes are unknown (Leppänen et al. 2018, Fu et al. 2020). The microbial composition can vary temporally in the air and, thus, air sampling generally represents a short-term microbial exposure. However, in the present study, the air sampling method provided an easy way to evaluate differences between schools.

In the characterization of the cultivable moulds in indoor airborne and settled dust, special attention was paid on mycoparasitic, mycotrophic *Trichoderma* species which were interestingly found only in problematic schools. The authors then propose that the presence of mycoparasitic species may serve as a bioindicator for the presence of indoor mold growth resulting from moisture damage. The role and significance of individual species, including potentially pathogenic *Trichoderma* species, on occupant exposure is out of the scope of this study. Considering the diversity of environmental microorganisms, it is suspected that the presence of single species cannot alone explain the perceived differences between buildings.

There are some limitations in our study regarding the number of studied buildings and classrooms. According to the authors, the investigation of six studied classrooms per school provided a representative picture about typical school conditions, since the classrooms in Finnish schools are notably clean and well-ventilated, and very similar to each other considering the construction and furniture materials or pollutants indoors. These school buildings well represent the modern Nordic schools with mechanical supply and exhaust heat recovery ventilation. Although the number of the studied schools was enough to fulfil the aim of this study, results and conclusions of the present study can be used in the future to investigate a higher number of problematic and comparison buildings. Finally, a longitudinal study on problematic and comparison schools, addressing few critical factors with a high seasonal variation and with simultaneous recording of occupants' perceptions, could provide useful information.

4. Conclusions

Altogether 14 different IAQ-related parameters were studied, and occupants' environmental perceptions and symptoms were recorded with two types of questionnaires. The teachers' and students' answers to the indoor climate questionnaires showed differences among symptom prevalence and perceived work environment and confirmed the initial categorisation of the schools (problematic and comparison). Symptom prevalence and perceived work environment varied even in nearby classrooms; thus, the developed and tested online questionnaire was an informative tool to assess perceived IAQ in a school building, supporting further development.

The concentrations of airborne cultivable fungi and identified fungal species in dust demonstrated most clearly the differences between schools in which occupants did or did not report IAQ-related problems. More local differences were found in the airborne microbial concentrations in the problematic schools. The results indicated a possible presence of indoor microbial sources in schools P2 and P3.

In the dust samples from exhaust air filters and classrooms, special attention was paid to pathogenic and/or toxigenic *Trichoderma* strains, and interestingly, they were found only in the problematic schools. Mycoparasitic and mycotrophic *T. atroviride* strains were the most frequently isolated *Trichoderma* species, which could indicate hidden excess moisture and mould growth of other fungi. Viable conidia of the potentially pathogenic species *T. longibrachiatum* and *T. citrinoviride* were also of concern in problematic school buildings; however, they

seemed to be more sensitive towards the tested biocides than *T. atroviride* and *T. trixiae*.

The concentration levels of the other studied pollutants were very low in all schools, with no differences between buildings. In other words, many of the measurement methods commonly used in indoor air investigations were unable to detect problems. Finding alternative methods needs more attention in future studies. Indicator methods for excess moisture and mould, such as exhaust air filter analyses in this study, need further study with simultaneous investigations of supply air filters and ventilation pressure conditions as potential contaminant transport mechanisms.

CRediT authorship contribution statement

Camilla Vornanen-Winqvist: Investigation, Data curation, Writing - original draft, Visualization. **Kati Järvi:** Formal analysis, Investigation, Writing - review & editing. **Maria A. Andersson:** Methodology, Validation, Investigation, Writing - review & editing, Visualization. **Caroline Duchaine:** Resources, Writing - review & editing. **Valérie Létourneau:** Validation, Investigation, Writing - review & editing. **Orsolya Kedves:** Investigation, Writing - review & editing. **László Kredics:** Validation, Investigation, Resources, Writing - review & editing. **Raimo Mikkola:** Investigation, Writing - review & editing. **Jarek Kurnitski:** Conceptualization, Writing - review & editing, Project administration, Funding acquisition. **Heidi Salonen:** Conceptualization, Data curation, Writing - review & editing, Supervision, Project administration, Funding acquisition.

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Author Contributions

Camilla Vornanen-Winqvist took the lead in writing the manuscript, participated in the study design and performed the experiments. Kati Järvi performed the experiments and analysed the online questionnaire data. Maria A. Andersson designed and performed the mycobiota experiments, analysed the data and participated in writing the paper. Raimo Mikkola contributed to measuring IAQ factors and analysed the VOC content of samples. Orsolya Kedves and László Kredics designed and performed the identification of the fungal morphotypes. Caroline Duchaine and Valérie Létourneau designed and performed the analyses of endotoxins and reviewed the manuscript. Jarek Kurnitski was responsible for designing the experiments. Heidi Salonen, the principal investigator, supervised the study design and contributed to the manuscript writing. All authors discussed the results and commented on the manuscript.

Conflicts of Interest

The authors declare no conflict of interest. The funding sponsors had no role in the design of the study; in the collection, analyses, or interpretation of data; in the writing of the manuscript, and in the

decision to publish the results.

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Appendix A. Supplementary material

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References

- Andersson, K., 1998. Epidemiological approach to indoor air problems. *Indoor Air* 8 (S4), 32–39.
- Annesi-Maesano, I., Baiz, N., Banerjee, S., Rudnai, P., Rive, S., SINPHONIE Group, 2013. Indoor air quality and sources in schools and related health effects. *J. Toxicol. Environ. Health, Part B* 16 (8), 491–550.
- Barnes, C., 2019. Fungi and atopy. *Clin. Rev. Allergy Immunol.* 57 (3), 439–448.
- Bartlett, K.H., Kennedy, S.M., Brauer, M., Van Netten, C., Dill, B., 2004. Evaluation and a predictive model of airborne fungal concentrations in school classrooms. *Ann. Occup. Hyg.* 48 (6), 547–554.
- Bennett, J., Davy, P., Trompetter, B., Wang, Y., Pierse, N., Boulic, M., Phipps, R., Howden-Chapman, P., 2019. Sources of indoor air pollution at a New Zealand urban primary school; a case study. *Atmos. Pollut. Res.* 10 (2), 435–444.
- Bornehag, C., Sundell, J., Sigsgaard, T., Janson, S., 2006. Potential self-selection bias in a nested case-control study on indoor environmental factors and their association with asthma and allergic symptoms among pre-school children. *Scand. J. Public Health* 34 (5), 534–543.
- Brauer, C., Mikkelsen, S., 2010. The influence of individual and contextual psychosocial work factors on the perception of the indoor environment at work: a multilevel analysis. *Int. Arch. Occup. Environ. Health* 83 (6), 639–651.
- Brauer, C., Mikkelsen, S., 2003. The context of a study influences the reporting of symptoms. *Int. Arch. Occup. Environ. Health* 76 (8), 621–624.
- Castagnoli, E., Marik, T., Mikkola, R., Kredics, L., Andersson, M.A., Salonen, H., Kurnitski, J., 2018. Indoor *Trichoderma* strains emitting peptaibols in guttation droplets. *J. Appl. Microbiol.* 125 (5), 1408–1422.
- Chatzidiakou, L., Mumovic, D., Summerfield, A.J., 2012. What do we know about indoor air quality in school classrooms? A critical review of the literature. *Intell. Build. Int.* 4 (4), 228–259.
- Chen, L., Jennison, B.L., Yang, W., Omaye, S.T., 2000. Elementary school absenteeism and air pollution. *Inhalation Toxicol.* 12 (11), 997–1016.
- Chithra, V., Nagendra, S., 2018. A review of scientific evidence on indoor air of school building: pollutants, sources, health effects and management. *Asian J. Atmos. Environ.* 12 (2), 87–108.
- Currie, J., Hanushek, E.A., Kahn, E.M., Neidell, M., Rivkin, S.G., 2009. Does pollution increase school absences? *Rev. Econ. Stat.* 91 (4), 682–694.
- Druzhinina, I.S., Seidl-Seiboth, V., Herrera-Estrella, A., Horwitz, B.A., Kenerley, C.M., Monte, E., Mukherjee, P.K., Zeilinger, S., Grigoriev, I.V., Kubicek, C.P., 2011. *Trichoderma*: the genomics of opportunistic success. *Nat. Rev. Microbiol.* 9 (10), 749–759.
- Duquenne, P., Marchand, G., Duchaine, C., 2013. Measurement of endotoxins in bioaerosols at workplace: a critical review of literature and a standardization issue. *Ann. Occup. Hygiene* 57 (2), 137–172.
- Ervasti, J., Kivimäki, M., Kawachi, I., Subramanian, S., Pentti, J., Oksanen, T., Puusniekka, R., Pohjonen, T., Vahtera, J., Virtanen, M., 2012. School environment as predictor of teacher sick leave: data-linked prospective cohort study. *BMC Public Health* 12 (1) pp. article No. 770.
- Fang, L., Wyon, D.P., Clausen, G., Fanger, P.O., 2004. Impact of indoor air temperature and humidity in an office on perceived air quality, SBS symptoms and performance. *Indoor Air* 14, 74–81.
- Faustman, E.M., Silbernagel, S.M., Fenske, R.A., Burbacher, T.M., Ponce, R.A., 2000. Mechanisms underlying children's susceptibility to environmental toxicants. *Environ. Health Perspect.* 108 (Suppl 1), 13–21.
- Finell, E., Haverinen-Shaughnessy, U., Tolvanen, A., Laaksonen, S., Karvonen, S., Sund, R., Saari, V., Luopa, P., Ståhl, T., Putus, T., 2017. The associations of indoor environment and psychosocial factors on the subjective evaluation of indoor air quality among lower secondary school students: a multilevel analysis. *Indoor Air* 27 (2), 329–337.
- Finell, E., Tolvanen, A., Haverinen-Shaughnessy, U., Laaksonen, S., Karvonen, S., Sund, R., Luopa, P., Pekkanen, J., Ståhl, T., 2018. Indoor air problems and the perceived social climate in schools: A multilevel structural equation analysis. *Sci. Total Environ.* 624, 1504–1512.
- Finnish Meteorological Institute, 2019-last update. Weather observation data. Available: <https://ilmatieteenlaitos.fi/havaintojen-lataus#!/> [6/16, 2019].
- FIOH, 2020-last update. The indoor climate questionnaire [Homepage of Finnish Institute of Occupational Health], [Online]. Available: <https://www.ttl.fi/palvelu/sisailmastokysely/> [01/20, 2020].
- FIOH, 2017-last update. Threshold values for indoor environment [In Finnish] [Homepage of Finnish Institute of Occupational Health], [Online]. Available: <https://www.ttl.fi/wp-content/uploads/2016/09/sisaympariston-viitearvoja.pdf> [6/16, 2019].
- Fu, X., Norbäck, D., Yuan, Q., Li, Y., Zhu, X., Hashim, J.H., Hashim, Z., Ali, F., Zheng, Y., Lai, X., 2020. Indoor microbiome, environmental characteristics and asthma among junior high school students in Johor Bahru Malaysia. *Environ. Int.* 138, 105664.
- Haverinen-Shaughnessy, U., Borrás-Santos, A., Turunen, M., Zock, J., Jacobs, J., Krop, E., Casas, L., Shaughnessy, R., Täubel, M., Heederik, D., 2012. Occurrence of moisture problems in schools in three countries from different climatic regions of Europe based on questionnaires and building inspections—the HITEA study. *Indoor Air* 22 (6), 457–466.
- Hellgren, U.M., 2008. Complaints and symptoms among hospital staff in relation to indoor air and the condition and need for repairs in hospital buildings. *Scand. J. Work Environ. Health* 34 (4), 58–63.
- Husman, T., Meklin, T., Vepsäläinen, A., Vahteristo, M., Koivisto, J., Halla-aho, J., Hyvärinen, A., Koponen, V., Nevalainen, A., 2002. Respiratory infections among children in moisture damaged schools. *Indoor Air* 2, 484–487.
- ISO 16000-6:2011. Indoor air — Part 6: Determination of volatile organic compounds in indoor and test chamber air by active sampling on Tenax TA sorbent, thermal desorption and gas chromatography using MS or MS-FID.
- Jacobs, J., Borrás-Santos, A., Krop, E., Täubel, M., Leppanen, H., Haverinen-Shaughnessy, U., Pekkanen, J., Hyvärinen, A., Doeke, G., Zock, J., Heederik, D., 2014a. Dampness, bacterial and fungal components in dust in primary schools and respiratory health in schoolchildren across Europe. *Occup. Environ. Med.* 71 (10), 704–712.
- Jacobs, J., Krop, E., Borrás-Santos, A., Zock, J., Täubel, M., Hyvärinen, A., Pekkanen, J., Doeke, G., Heederik, D., HITEA Schools Study Consortium, 2014b. Endotoxin levels in settled airborne dust in European schools: the HITEA school study. *Indoor Air* 24 (2), 148–157.
- Jacobs, J., Krop, E., de Wind, S., Spithoven, J., Heederik, D., 2013. Endotoxin levels in homes and classrooms of Dutch school children and respiratory health. *Eur. Respir. J.* 42 (2), 314–322.
- Järvi, K., Vornanen-Winqvist, C., Mikkola, R., Kurnitski, J., Salonen, H., 2018. Online questionnaire as a tool to assess symptoms and perceived indoor air quality in a school environment. *Atmosphere* 9 (7) article No. 270.
- Kanchongkittiphon, W., Mendell, M., Gaffin, J., Wang, G., Phipatanakul, W., 2015. Indoor environmental exposures and exacerbation of asthma: an update to the 2000 review by the Institute of Medicine. *Environ. Health Perspect.* 123 (1), 6–20.
- Kudo, E., Song, E., Yockey, L.J., Rakib, T., Wong, P.W., Homer, R.J., Iwasaki, A., 2019. Low ambient humidity impairs barrier function and innate resistance against influenza infection. *Proc. Natl. Acad. Sci. U. S. A.* 116 (22), 10905–10910.
- Lahtinen, M., Sundman-Digert, C., Reijula, K., 2004. Psychosocial work environment and indoor air problems: a questionnaire as a means of problem diagnosis. *Occup. Environ. Med.* 61 (2), 143–149.
- Lahtinen, M., Huuhtanen, P., Kähkönen, E., Reijula, K., 2002. Psychosocial dimensions of solving an indoor air problem. *Indoor Air* 12 (1), 33–46.
- Lai, P.S., Sheehan, W.J., Gaffin, J.M., Petty, C.R., Coull, B.A., Gold, D.R., Phipatanakul, W., 2015. School endotoxin exposure and asthma morbidity in inner-city children. *Chest* 148 (5), 1251–1258.
- Lin, L., Wargocki, P., Wyon, D.P., Lian, Z., 2011. Effects of thermal discomfort in an office on perceived air quality, SBS symptoms, physiological responses, and human performance. *Indoor Air* 21 (5), 376–390.
- Leppänen, H.K., Täubel, M., Jayaprakash, B., Vepsäläinen, A., Pasanen, P., Hyvärinen, A., 2018. Quantitative assessment of microbes from samples of indoor air and dust. *J. Exposure Sci. Environ. Epidemiol.* 28 (3), 231.
- Lesire, V., Hazouard, E., Dequin, P., Delain, M., Therizol-Ferly, M., Legras, A., 1999. Possible role of *Chaetomium globosum* in infection after autologous bone marrow transplantation. *Intensive Care Med.* 25 (1), 124–125.
- Letourneau, V., Nehme, B., Meriaux, A., Masse, D., Duchaine, C., 2009. Impact of production systems on swine confinement buildings bioaerosols. *J. Occupat. Environ. Hygiene* 7 (2), 94–102.
- Lowen, A.C., Mubareka, S., Steel, J., Palese, P., 2007. Influenza virus transmission is dependent on relative humidity and temperature. *PLoS Pathogens* 3 (10) article No. e151.
- Madureira, J., Alvim-Ferraz, M., Rodrigues, S., Gonçalves, C., Azevedo, M., Pinto, E., Mayan, O., 2009. Indoor air quality in schools and health symptoms among Portuguese teachers. *Hum. Ecol. Risk Assess.* 15 (1), 159–169.
- Madureira, J., Paciência, I., Oliveira Fernandes, E., 2012. Levels and indoor-outdoor relationships of size-specific particulate matter in naturally ventilated Portuguese schools. *J. Toxicol. Environ. Health, Part A* 75 (22–23), 1423–1436.
- Madureira, J., Paciência, I., Rufo, J., Ramos, E., Barros, H., Teixeira, J., Oliveira Fernandes, E., 2015. Indoor air quality in schools and its relationship with children's respiratory symptoms. *Atmos. Environ.* 118, 145–156.
- Meklin, T., 2002. Microbial exposure and health in schools-effects of moisture damage and renovation. Ph.D. dissertation, National Public Health Institute.
- Meklin, T., Putus, T., Pekkanen, J., Hyvärinen, A., Hirvonen, M., Nevalainen, A., 2005. Effects of moisture-damage repairs on microbial exposure and symptoms in schoolchildren. *Indoor Air* 15 (Suppl. 10), 40–47.
- Mendell, M.J., Heath, G.A., 2005. Do indoor pollutants and thermal conditions in schools influence student performance? A critical review of the literature. *Indoor Air* 15 (1), 27–52.
- Mendell, M.J., Mirer, A., 2009. Indoor thermal factors and symptoms in office workers: findings from the US EPA BASE study. *Indoor Air* 19 (4), 291–302.
- Mikkola, R., Andersson, M.A., Hautaniemi, M., Salkinoja-Salonen, M.S., 2015. Toxic indole alkaloids avrainvillamide and toxaphen B produced by a biocide tolerant indoor mold *Aspergillus westerdijkiae*. *Staphen* 99, 58–67.
- Ministry of Social Affairs and Health, Finland, 2015. Decree of the Ministry of Social

- Affairs and Health on health-related conditions of housing and other residential buildings and qualification requirements for third-party experts (545/2015). Available at: <https://www.finlex.fi/en/laki/kaannokset/2015/en20150545.pdf>. Finland.
- Ministry of the Environment, Finland, 2017. Decree of the Ministry of the Environment on indoor climate and ventilation of a new building (Ympäristöministeriön asetus uuden rakennuksen sisäilmastosta ja ilmanvaihdosta 1009/2017). Available in Finnish: <https://www.finlex.fi/fi/laki/alkup/2017/20171009#Pidp446795120>.
- Mishra, N., Bartsch, J., Ayoko, G.A., Salthammer, T., Morawska, L., 2015. Volatile organic compounds: characteristics, distribution and sources in urban schools. *Atmos. Environ.* 106, 485–491.
- Nissilä, J., Savelieva, K., Lampi, J., Ung-Lanki, S., Elovainio, M., Pekkanen, J., 2019. Parental worry about indoor air quality and student symptom reporting in primary schools with or without indoor air quality problems. *Indoor Air* 29 (5), 865–873.
- Noss, I., Wouters, I.M., Visser, M., Heederik, D.J., Thorne, P.S., Brunekreef, B., Doekes, G., 2008. Evaluation of a low-cost electrostatic dust fall collector for indoor air endotoxin exposure assessment. *Appl. Environ. Microbiol.* 74 (18), 5621–5627.
- Park, J.H., Cox-Ganser, J., Rao, C., Kreiss, K., 2006. Fungal and endotoxin measurements in dust associated with respiratory symptoms in a water-damaged office building. *Indoor Air* 16 (3), 192–203.
- Pasanen, T., Keskikuru, T., Ridell, J., Pasanen, P., 2002. Indoor air quality and the adequacy of cleaning in 25 Finnish schools. In: Levin, H. (Ed.), *Indoor Air, 2002. Proceedings: 9th International Conference on Indoor Air Quality and Climate, Monterey, California, June 30 - July 5, 2002*, p. 87–91. Santa Cruz, California: Indoor Air 2002.
- Peitzsch, M., Sulyok, M., Täubel, M., Vishwanath, V., Krop, E., Borrás-Santos, A., Hyvärinen, A., Nevalainen, A., Krska, R., Larsson, L., 2012. Microbial secondary metabolites in school buildings inspected for moisture damage in Finland, The Netherlands and Spain. *J. Environ. Monit.* 14 (8), 2044–2053.
- Pyrri, I., Kapsanaki-Gotsi, E., 2017. Functional relations of airborne fungi to meteorological and pollution factors in a Mediterranean urban environment. *Fungal Ecol.* 30, 48–54.
- Reijula, K., Sundman-Digert, C., 2004. Assessment of indoor air problems at work with a questionnaire. *Occup. Environ. Med.* 61 (1), 33–38.
- Ruggieri, S., Longo, V., Perrino, C., Canepari, S., Drago, G., L'Abbate, L., Balzan, M., Cuttitta, G., Scaccianoce, G., Minardi, R., 2019. Indoor air quality in schools of a highly polluted south Mediterranean area. *Indoor Air* 29 (2), 276–290.
- Runeson-Broberg, R., Norbäck, D., 2013. Sick building syndrome (SBS) and sick house syndrome (SHS) in relation to psychosocial stress at work in the Swedish workforce. *Int. Arch. Occup. Environ. Health* 86 (8), 915–922.
- Salonen, H., Duchaine, C., Létourneau, V., Mazaheri, M., Clifford, S., Morawska, L., 2013. Endotoxins in indoor air and settled dust in primary schools in a subtropical climate. *Environ. Sci. Technol.* 47 (17), 9882–9890.
- Salonen, H., Duchaine, C., Létourneau, V., Mazaheri, M., Laitinen, S., Clifford, S., Mikkola, R., Lappalainen, S., Reijula, K., Morawska, L., 2016. Endotoxin levels and contribution factors of endotoxins in resident, school, and office environments – a review. *Atmos. Environ.* 142, 360–369.
- Salonen, H., Duchaine, C., Mazaheri, M., Clifford, S., Lappalainen, S., Reijula, K., Morawska, L., 2015. Airborne viable fungi in school environments in different climatic regions – a review. *Atmos. Environ.* 104, 186–194.
- Salthammer, T., Uhde, E., Schripp, T., Schieweck, A., Morawska, L., Mazaheri, M., Clifford, S., He, C., Buonanno, G., Querol, X., 2016. Children's well-being at schools: Impact of climatic conditions and air pollution. *Environ. Int.* 94, 196–210.
- Samson, R.A., Hoekstra, E.S., Frisvad, J.C., 2004. Introduction to food- and airborne fungi. Centraalbureau voor Schimmelcultures (CBS).
- Savelieva, K., Marttila, T., Lampi, J., Ung-Lanki, S., Elovainio, M., Pekkanen, J., 2019. Associations between indoor environmental quality in schools and symptom reporting in pupil-administered questionnaires. *Environ. Health* 18 (1), 1–12.
- Seidl, V., Song, L., Lindquist, E., Gruber, S., Koptchinskiy, A., Zeilinger, S., Schmoll, M., Martínez, P., Sun, J., Grigoriev, I., 2009. Transcriptomic response of the mycoparasitic fungus *Trichoderma atroviride* to the presence of a fungal prey. *BMC Genom.* 10 (1) pp. article No. 567.
- Sheehan, W.J., Hoffman, E.B., Fu, C., Baxi, S.N., Bailey, A., King, E., Chapman, M.D., Lane, J.P., Gaffin, J.M., Permaul, P., 2012. Endotoxin exposure in inner-city schools and homes of children with asthma. *Ann. Allergy Asthma Immunol.* 108 (6), 418–422.
- Śmiełowska, M., Marć, M., Zabiegała, B., 2017. Indoor air quality in public utility environments - A review. *Environ. Sci. Pollut. Res.* 24 (12), 11166–11176.
- Spaan, S., Heederik, D., Thorne, P., Wouters, I., 2007. Optimization of airborne endotoxin exposure assessment: effects of filter type, transport conditions, extraction solutions, and storage of samples and extracts. *Appl. Environ. Microbiol.* 73 (19), 6134–6143.
- Spaan, S., Doekes, G., Heederik, D., Thorne, P., Wouters, I., 2008. Effect of extraction and assay media on analysis of airborne endotoxin. *Appl. Environ. Microbiol.* 74 (12), 3804–3811.
- Sunwoo, Y., Chou, C., Takeshita, J., Murakami, M., Tochihara, Y., 2006. Physiological and subjective responses to low relative humidity. *J. Physiol. Anthropol.* 25 (1), 7–14.
- Tähtinen, K., Lappalainen, S., Karvala, K., Lahtinen, M., Salonen, H., 2019. Probability of Abnormal indoor air exposure categories compared with occupants' symptoms, health information, and psychosocial work environment. *Appl. Sci.* 9 (1), 99.
- Theodosiou, T., Ordoumpozanis, K., 2008. Energy, comfort and indoor air quality in nursery and elementary school buildings in the cold climatic zone of Greece. *Energy Build.* 40 (12), 2207–2214.
- Torresin, S., Pernigotto, G., Cappelletti, F., Gasparella, A., 2018. Combined effects of environmental factors on human perception and objective performance: A review of experimental laboratory works. *Indoor Air* 28 (4), 525–538.
- Turunen, M., Toyinbo, O., Putus, T., Nevalainen, A., Shaughnessy, R., Haverinen-Shaughnessy, U., 2014. Indoor environmental quality in school buildings, and the health and wellbeing of students. *Int. J. Hyg. Environ. Health* 217 (7), 733–739.
- Vornanen-Winqvist, C., Järvi, K., Toomla, S., Ahmed, K., Andersson, M.A., Mikkola, R., Marik, T., Kredics, L., Salonen, H., Kurnitski, J., 2018a. Ventilation positive pressure intervention effect on indoor air quality in a school building with moisture problems. *Int. J. Environ. Res. Public Health* 15 (2) article No. 230.
- Vornanen-Winqvist, C., Salonen, H., Järvi, K., Andersson, M., Mikkola, R., Marik, T., Kredics, L., Kurnitski, J., 2018b. Effects of ventilation improvement on measured and perceived indoor air quality in a school building with a hybrid ventilation system. *Int. J. Environ. Res. Public Health* 15 (7) article No. 1414.
- WHO, 2010. WHO Guidelines for Indoor Air Quality: Selected Pollutants. World Health Organization, Regional Office Europe.
- WHO, 2009. WHO Guidelines for Indoor Air Quality: Dampness and Mould. World Health Organization, Regional Office Europe.
- WHO, 2006. Principles for Evaluating Health Risks in Children Associated with Exposure to Chemicals. World Health Organization.
- Wolkoff, P., 2018. Indoor air humidity, air quality, and health: An overview. *Int. J. Hyg. Environ. Health* 221 (3), 376–390.
- Wolkoff, P., Kjergaard, S.K., 2007. The dichotomy of relative humidity on indoor air quality. *Environ. Int.* 33 (6), 850–857.
- Yassin, M., Pillai, A., 2018. Monitoring of volatile organic compounds in different schools: a determinant of the indoor air quality. *Int. J. Environ. Sci. Technol.* 16 (6), 2733–2744.
- Zhong, L., Su, F., Batterman, S., 2017. Volatile organic compounds (VOCs) in conventional and high performance school buildings in the US. *Int. J. Environ. Res. Public Health* 14 (1) article No. 100.